Is there tree senescence? The fecundity evidence

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Despite its importance for forest regeneration, food webs, and human economies, changes in tree fecundity with tree size and age remain largely unknown. The allometric increase with tree diameter assumed in ecological models would substantially overestimate seed contributions from large trees if fecundity eventually declines with size. Current estimates are dominated by overrepresentation of small trees in regression models. We combined global fecundity data, including a substantial representation of large trees. We compared size–fecundity relationships against traditional allometric scaling with diameter and two models based on crown architecture. All allometric models fail to describe the declining rate of increase in fecundity with diameter found for 80% of 597 species in our analysis. The strong evidence of declining fecundity, beyond what can be explained by crown architectural change, is consistent with physiological decline. A downward revision of projected fecundity of large trees can improve the next generation of forest dynamic models.

tree fecundity | tree senescence | tree life history | allometric scaling | crown architecture

“Belgium, Luxembourg, and The Netherlands are characterized by “young” apple orchards, where over 60% of the trees are under 10 years old. In comparison, Estonia and the Czech Republic have relatively “old” orchard[s] with almost 60% and 43% over 25 years old” (1).
Significance

Physiological decline in large, old trees could involve diminished reproductive effort, but evidence is lacking. A global analysis finds that fecundity decline in large trees is pervasive, consistent with physiological decline, thus providing new evidence for tree senescence. For the 20% of species not showing fecundity declines, this lack of evidence was linked not to specific species groups, but rather to lack of large trees in the data. Large trees remain important for their structural attributes, but results highlight the critical role of intermediate trees for their contribution to reproduction.

Unfortunately, underrepresentation of big trees in forests frustrates efforts to infer how fecundity changes with size. Simple allometric relationships between seed production and tree diameter can offer useful predictions for the small- to intermediate-size trees that dominate observational data, so it is not surprising that modeling began with the assumption of allometric scaling (32–36). Extrapolation from these models would predict that seed production by the small trees from which most observations come may be overwhelmed by big trees. Despite the increase with tree size assumed by ecologists (37), evidence for declining reproduction in large trees has continued to accumulate from horticultural practice (3, 4, 38, 39) and at least some ecological (40–45) and forestry literature (46, 47). However, we are unaware of studies that evaluate changes in fecundity that include substantial numbers of large trees.

Understanding the role of size and age is further complicated by the fact that tree fecundity ranges over orders of magnitude from tree to tree of the same species and within the same tree from year to year—a phenomenon known as “masting.” The variation in seed-production data requires large sample sizes not only to infer the effects of size, but also to account for local habitat and interannual climate variation. For example, a one-time destructive harvest to count seeds in felled trees (48, 49) misses the fact that the same trees would offer a different picture had they been harvested in a different year. An oak that produces 100 acorns this year may produce 10,000 next year. A pine that produces 500 cones this year can produce zero next year. Few datasets offer the sample sizes of trees and tree years needed to estimate effects of size and habitat conditions in the face of this high intertree and interyear variability (43).

We begin this analysis by extending allometric scaling to better reflect the geometry of fecundity with tree size. We then reexamine the size–fecundity relationship using data from the Mast- ing Inference and Forecasting (MASTIF) project (50), which includes substantial representation of large trees, and a modeling framework that allows for the possibility that fecundity plateaus or even declines in large trees. Unlike previous studies, we account for the nonallometric influences that come through competition and climate. We demonstrate that fecundity–diameter relationships depart substantially from allometric scaling in ways that are consistent with physiological senescence.

Continuous increase with size has been assumed in most models of tree fecundity, supported in part by allometric regressions against diameter, typically of the form

$$\log M_f = \beta_0 + \beta_D \log D$$

for fecundity mass \(M_f = m \times f\) (48, 51), where \(D\) is tree diameter, \(m\) is mass per seed, and fecundity \(f\) is seeds per tree per year. Of course, this model cannot be used to determine whether or how fecundity changes with tree diameter unless expanded to include additional quadratic or higher-order terms (52).

The assumption of continual increase in fecundity was interpreted from early seed-trap studies, which initially assumed that \(\beta_D = 2\), i.e., fecundity proportional to stem basal area (33–34, 51). Models subsequently became more flexible, first with \(\beta_D\) values fitted, rather than fixed, yielding estimates in the range (0.3, 0.9) in one study (ref. 52, 18 species) and (0, 4.1) in another (ref. 56, 4 species). However, underrepresentation of large trees in typical datasets means that model fitting is dominated by the abundant small size classes.

To understand why data and models could fail to accurately represent change in fecundity with size, consider that allometric scaling in Eq. 1 can be maintained dynamically only if change in both adheres to a strict proportionality

$$\frac{1}{f} \frac{df}{dt} \times \frac{1}{D} \frac{dD}{dt} = \text{constant}$$
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For allometric scaling, any variable that affects diameter growth has to simultaneously affect change in fecundity and in the same, proportionate way. In other words, allometric scaling cannot hold if there are selective forces on fecundity that do not operate through diameter growth and vice versa.

On top of this awkward constraint that demands proportionate responses of growth and fecundity, consider further that standard arguments for allometric scaling are not directly relevant for tree fecundity. Allometry is invoked for traits that maintain relationships between body parts as an organism changes size (29). For example, a diameter increment translates to an increase in volume throughout the tree (58, 59). Because the cambial layer essentially blankets the tree, a volume increment cannot depart much from a simple allometric relationship with diameter. However, the same cannot be said for all plant parts, many of which clearly do not allometrically scale; for example, seed size does not scale with leaf size (60), presumably because structural constraints are not the dominant forces that relate them (61).

To highlight why selective forces might not generate strict allometric scaling for reproduction, consider that a tree allocates a small fraction of potential buds to reproduction in a given year (62, 63). Still, if the number of buds on a tree bears some direct relationship to crown dimensions and, thus, diameter, there might be allometric scaling. However, the fraction of buds allocated to reproduction and their subsequent development to seed is affected by interannual weather and other selective forces (e.g., bud abortion, pollen limitation) in ways that diameter growth is not (64–66). In fact, weather might have opposing effects on growth and reproduction (67). Furthermore, resources can change the relationship between diameter and fecundity, including light levels (52, 68–70) and atmospheric CO2 (71).

Some arguments based on carbon balance anticipate a decline in fecundity with tree size (72). Increased stomatal limitation (11) and reduced leaf turgor pressure (14, 73) from increasing hydraulic path length could reduce carbon gains in large trees. Assimilation rates on a leaf area basis can decline with tree size (74), while respiration rate per leaf area can increase [Sequoia sempervirens (75), Liquidambar styraciflua (76), and Pinus sylvestris (77)], consistent with the notion that whole-plant respiration rate may roughly scale with biomass (78). Maintenance respiration costs scale with diameter and in some typical species (79) but perhaps not in Pinus contorta and Picea engelmannii (80). Self-pruning of lower branches can reduce maintenance costs (81), but the ratio of carbon gain to respiration cost can still decline with size, especially where leaf area plateaus and per-area assimilation rates of leaves decline in large trees.

The question of size–fecundity relationships is related indirectly to the large literature on interannual variation in growth–fecundity allocation (3, 4, 43, 67, 82–87). The frequency and timing of mast years and species differences in the volatility of seed production can be related to short-term changes in physiological state and pollen limitation that might not predict the long-term relationships between size and reproductive effort. The interannual covariance in diameter growth and reproductive effort can range from strong in some species to weak in others (70, 87, 88). Understanding the relationships between short-term allocation and size–fecundity differences will be an important focus of future research.

Estimating effects of size on fecundity depends on the distribution of diameter data, \([D]\), where the bracket notation indicates a distribution or density. For some early-successional species, the size distribution changes from dominance by small trees in young stands to absence of small trees in old stands. If our goal was to describe the population represented by a forest inventory plot, we would typically think about the joint distribution of fecundity and diameter values, \([f, D] = [fD][D]\), that is represented by the sample. The size–fecundity relationship estimated for a stand at different successional stages would diverge simply due to the distribution of diameters, i.e., differences in \([D]\). For example, application of Eq. 1 to harvested trees selected to balance size classes (uniform \([D]\)) (48) overpredicts fecundity for large trees (49), but the relevance of such regressions for natural stands, where large trees are often rare, is unclear. Studies that expand Eq. 1 to allow for changing relationships with tree size now provide increasing evidence for a departure from allometric scaling in large trees (43, 70), despite dominance by small- to intermediate-size trees in these datasets. Here our goal is to understand the size–fecundity relationship \([fD]\) as an attribute of a species, i.e., not tied to a specific distribution of size classes observed in the particular stand studied.

The well-known weak relationship between tree size and age that comes from variable growth histories makes it important to clarify the implications of any finding of fecundity that declines with tree size: Can it happen if there are not also fecundity declines with tree age? The only argument for continuing increase in fecundity with age in the face of observed decreases with size would have to assume that the biggest trees are also the youngest trees. Of course, a large individual can be younger than a small individual. However, at the species level, integrating over populations sampled widely, mean diameter increases with age; at the species level, declines with size also imply declines with age. Estimating accurate species-level size effects requires distributed data and large sample sizes. The analysis here fits species-level parameters, with 585,670 trees and 10,542,239 tree years across 597 species.

Phylogenetic analysis might provide insight into the pervasiveness of fecundity declines with size. Inferring change in fecundity with size necessarily requires more information than is needed to fit a single slope parameter \(\beta_0\) in the simple allometric model. The noisier the data, the more difficult it becomes to estimate the additional parameters that are needed to describe changes in the fecundity relationship with size. We thus expect that noise alone will preclude finding size-related change in some species, depending on sample size and non–size-related variation. If the vagaries of noisy data and the distribution of diameters preclude estimation of declines in some species, then we do not expect that phylogeny will explain which species do and do not show these declines. Rather than phylogeny, this explanation would instead be tied to sample size and the distribution of diameter data. Conversely, phylogenetic conservatism, i.e., a tendency for declines to be clustered in related species, could suggest that fecundity declines are real.

To understand how seed production changes with tree size, our approach combines theory and data to evaluate allometric scaling and the alternative that fecundity may decline in large trees, consistent with physiological decline and senescence. We exploit two advances that are needed to determine how fecundity scales with tree size. First, datasets are needed with large trees, because studies in the literature often include few or none (85, 89, 90). Second, methods are introduced that are flexible to the possibility that fecundity continues to increase with size or not. We begin with a reformulation of allometric scaling, recognizing that change in fecundity could be regulated by size, without taking the form of Eq. 1 (Materials and Methods and SI Appendix, section S2). In other words, there could be allometric scaling with diameter, but it is not the relationship that has been used for structural quantities like biomass. We then analyze the relationships in data using a model that not only allows for potential changes in fecundity with size, but at the same time accounts for self-shading and shading by neighbors and for environmental variables that can affect fecundity and growth (Materials and Methods and SI Appendix, section S3). The fitted model is compared with our expanded allometric model to identify potential agreement. Finally, we examined phylogenetic trends in the species that do and do not show declines.
Results

Sixty-three percent of the 597 species in this study show eventual declines in fecundity with increasing diameter (type A in Materials and Methods) (Fig. 1 and Table 1). Seventy percent showed diminishing rates of increase (type B), indicating that fecundity might eventually plateau or even decrease. Only 20% show the continuous increase in fecundity (type C) that would be consistent with traditional allometric scaling.

The inflection in the fecundity–diameter relationship at $D_{opt}$ is species specific and exhibits phylogenetic coherence ($\lambda = 0.77$, $p < 10^{-25}$, $n = 281$) (Fig. 2). For the species with fecundity declines, $D_{opt}$ (the diameter where maximum fecundity occurs) averaged $41 \pm 35$ cm. We also compared $D_{opt}/D_{max}$ between species and it shows trends across phylogeny similar to those of $D_{opt}$ (SI Appendix, Fig. S5). Whereas phylogeny contributes to species differences in $D_{opt}$, the question of whether or not an optimum could be identified was not related to phylogeny (SI Appendix, Fig. S3: $\lambda < 10^{-13}$, $p \approx 1$, $n = 398$). Nor did we find a consistent pattern in the relationship between $D_{max}$ and the three types of fecundity change with diameter across genera from the temperate regions of North America, Europe, and Asia (SI Appendix, Fig. S4). For example, fecundity eventually declines in some Pinus species having large $D_{max}$ (e.g., Pinus monticola, Pinus lambertiana, and Pinus ponderosa), while declines could not be identified in Pinus jeffreyi.

Whether or not a decline can be detected was related not to phylogeny, but rather to the fraction of large trees in the data. We found that species estimated to have continuous increases in fecundity tend to have smaller proportions of large trees (SI Appendix, Fig. S6). In addition, the credible intervals were generally wider (higher transparencies in the lines of Fig. 1) in large size classes for species with fewer observations, as expected given the requirement that estimating declines at large size requires more diameter information (two parameters to estimate rather than one). A lack of large individuals can increase the uncertainties in our fecundity estimates, especially at large size.

Neither allometric model we examined (Materials and Methods) could explain the slowing and eventual decline in fecundity with tree size. Recall that we are not attempting to find a “best” model for tree allometry, but rather to determine whether alternative assumptions about crown surface could produce declines with tree size. Under assumptions of allometric scaling, crown surface area ($C_{SA}$) of open-grown trees tends to increase at an increasing rate with diameter; dots representing six species in Fig. 3A show no hint of eventual declining rates, despite flattening crowns (Fig. 4). Similar patterns were found under simulated shading conditions (SI Appendix, Fig. S2). Nor did we observe declines for 110 species in the ideal tree distribution (ITD) model (Materials and Methods) in Fig. 3B, which does not admit crown flattening, but does attempt to include idealized shading from neighbors.

Discussion

The fact that 80% of the 597 species tested here show declining rates of increase in fecundity with diameter (types A and B in Materials and Methods), and 63% of the total actually decrease...
Table 1. Percentage of species in three fecundity–diameter response classes (Materials and Methods) by temperate and tropical regions

<table>
<thead>
<tr>
<th>Region (no. species)</th>
<th>Sample size (tree years)</th>
<th>Type A: Eventual decline</th>
<th>Type B: Sigmoid increase</th>
<th>Type C: Continuous increase</th>
</tr>
</thead>
<tbody>
<tr>
<td>Temperate (169)</td>
<td>2,483,125</td>
<td>61.5</td>
<td>15.4</td>
<td>23.1</td>
</tr>
<tr>
<td>Tropical (428)</td>
<td>8,059,114</td>
<td>64.0</td>
<td>17.1</td>
<td>18.9</td>
</tr>
</tbody>
</table>

(type A), provides empirical evidence for declining fecundity with size (Fig. 1). Declining fecundity does not fit the standard allometric model based on stem diameter of Eq. 1. Nor can it be explained by allometric relationships with crown area (Fig. 4), by either of two models (Fig. 3). In other words, size-related fecundity declines do not align with architectural changes related to flattening crowns.

If declining fecundity is a defining feature of big trees, then why did we not identify it in all species? The parsimonious explanation is that the combination of overrepresentation of small trees and the added information requirement raises the bar for finding declines in large trees. Finding eventual declines in fecundity requires more information than is needed to estimate a single parameter for a rate of increase; the capacity to identify declines with size is expected to be greatest for species represented by a high proportion of large trees (SI Appendix, Fig. S6). If continuously increasing fecundity were associated with specific families or genera, then there would be cause for speculation on...
the genetic differences that lead to declines in some species, but not others. Absence of a phylogenetic pattern to explain which species do and do not decrease might be explained by the fact that only some species suffer physiological decline with size and age and that this immunity to aging has arisen multiple times in unrelated species. Alternatively, it is also consistent with the knowledge that accurate estimates depend importantly on the distribution of diameter data, and this relationship to the availability of large trees is shown here. We cannot offer a specific rule for identifying fecundity decline, because that depends on all sources of uncertainty in the data and model. Lack of a phylogenetic signal in whether or not we can estimate a decline, combined with few large trees for nondeclining species, suggests that fecundity decline may be more pervasive than estimated here.

Our results indicate that trees produce proportionately fewer seeds per diameter and per crown area as they increase in size at the species level, consistent with age-related decline. This support for horticultural evidence (3, 4, 38) is also consistent with studies suggesting plateaus and/or declines in fecundity for species of *Quercus* (46, 47, 92), *P. sylvestris* (93), and *Avicennia marina* (92). It confirms increasing reports of fecundity declines in large and old trees in tropical (42) and temperate regions (23, 40, 43, 44, 46, 47). It may agree with a carbon-balance perspective (94) and with a recent meta-analysis reporting that more than half of 20 woody perennial species exhibited age-related declines in reproduction, with only one species suggesting an increase with age (95).

The fact that allometric scaling with diameter fails for fecundity means that fecundity also does not scale with other

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**Fig. 3.** Neither allometric model can generate fecundity decline with size for realistic parameter ranges. (A) Crown surface area evaluated from images of six species (SI Appendix, Fig. S1). (B) Crown surface area evaluated from 110 species in North America using the ideal tree distribution (ITD) model (91). Colors indicate genera (ordered alphabetically for 13 main genera). Crown surface area C_{SA} continues to increase with diameter from observed and simulated data with C_{SA} for each species displayed as a proportion of the maximum C_{SA} value.

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The fact that allometric scaling with diameter fails for fecundity means that fecundity also does not scale with other
dimensional properties that scale with diameter, including basal area, biomass, or wood volume. For example, basal area $B$ scales with diameter $B \propto D^2$, as does biomass $M \propto D^3$, where coefficient $c$ varies by species. Because $B$ and $M$ both scale with diameter, they also scale with each other as a simple ratio of coefficients $M \propto B^{3/2}$. The converse is also true: Because fecundity does not scale with diameter, it cannot scale with basal area or biomass. Our results indicate that fecundity might only roughly be evaluated as a proportionality with stand basal area or biomass.

Our results do not separate the effects of age and size, but they add an additional dimension to existing evidence for aging, that of declining reproductive effort (Fig. 1). Age-related reductions in maximum photosynthetic capacity (72, 96), xylem growth rate (97), specific leaf area (21), and reproductive development (22) may combine with fecundity decline as part of an aging syndrome. Size alone may explain some of these changes if hydraulic limitation develops in tall trees (8, 11, 98, 99). Granting experiments can help to separate effects of size and age, at least some of which already provide evidence of senescence (7).

The finding that fecundity declines with size means that proportionate allocation to reproduction must also decline. Again, continuing increase in biomass increment is consistent with expanding cambial tissue, as evident from empirical data (58). Unlike biomass, fecundity is not structurally related to size, as bigger trees need not produce bigger seeds or even more of them. Foregoing current for future reproduction can pay when future benefits offset mortality risks (100–102). The exposure to selection on genes that confer high reproductive effort late in life can be essentially zero; this decline in the efficacy of selection for traits that could sufficiently boost later reproductive success would have to offset not only declining survival (103), but also the effects of accumulated deleterious mutations (104) and antagonistic pleiotropy (105, 106). Revised fecundity schedules available here can be combined with survival scenarios to update estimates of tree fitness.

The value of broad representation of species and size classes, combined with a methodology that reduces the dependence on the distribution of size classes, provides an additional perspective on basic dimensional relationships with fecundity. The massive numbers of observations in MASTIF distributed across climate and habitat space allowed us to estimate change in the diameter–fecundity relationship across the diameter range, while accounting for habitat variation. The capacity to combine the evidence from seed-trap studies with crop counts in the MASTIF model (44) played an important role, because seed-trap studies may not include large trees, whereas crop counts can selectively represent the important range of tree attributes. Conversely, small-seeded species that are poorly represented in crop-count studies are those for which fecundity estimates are most accurate in seed-trap studies (50).

The importance of fruit, seed, and nut production for human and animal consumption makes these results relevant to issues from food security to ecological food webs. The generality of fecundity declines with size reported here squares with patterns long recognized in orchard management and suggests the value of stronger connections between basic and applied research. The fact that fecundity is optimal or plateau at intermediate size is not just important for human fruit consumption, but also has implications for forest food webs. The large numbers of mast consumers in forests are affected by seed supply, especially as a winter food source (107, 108).

Despite the fact that fecundity directly modulates forest demographic processes (52, 109) and range shifts (110), insights here can improve on the assumed constant seed pool in Earth System models (111) and allometric scaling in stand simulators (112) that are used to examine effects of climate change (113).

Materials and Methods

Fecundity Data. Fecundity data were obtained from the MASTIF project (50) on 585,670 trees and 10,542,239 tree years from 597 species in five continents, including North America, South and Central America, Europe, Africa, and Asia (SI Appendix, Table S2). MASTIF data are restricted to seed traps and crop counts, both of which are referenced to fecundity in individual tree years (SI Appendix, section S1), supplemented by observations of maturation status in some datasets (52, 114). Crop counts refer to the current season’s production; we do not use counts of accumulated reproductive effort from species where cones might reside on trees for multiple years (e.g., ref. 115). Data from crop counts, seeds traps, and maturation statuses are detailed in previous publications (43, 44).

To gauge whether there might be insufficient numbers of large trees, we compared the distribution of diameters in MASTIF with the largest trees observed in tree inventory data, which were available for North America, Europe, and Asia (SI Appendix, Table S3). We evaluated a “near-maximum” diameter for each species $D_{\text{max}}$ to be the fifth largest tree in forest inventories. We used this order statistic, because, unlike quantities, it does not depend on the distribution of small trees in the data. We did not use the extreme value due to the potential for error in a single value (as opposed to five values). Due to limited inventory data relative to the numbers of species in South America and Africa, we converted maximum tree heights from Liu et al. (116) to diameter using allometric equations (117) (SI Appendix, Table S3). Approximately 11%, 7, and 3% of species had no trees larger than 40% of the $D_{\text{max}}$ in eastern North America, western North America, and Asia, respectively. For South and Central America, ~7% of species were less than 30% of the $D_{\text{max}}$. They were removed from further analysis. Five species that belong to the family Arecaceae were also excluded because diameter growth was irrelevant to fecundity for those palm species. There were 585,670 individual trees and 10,542,239 tree years in the remaining 597 species.

Crown Allometry. We cannot reject allometric scaling solely on the basis of poor fit to Eq. 1, because fecundity might scale with diameter in a different way. To fully evaluate allometric equations, including whether or not alternative assumptions could explain declines with size, we examined scaling with crown surface area and crown depth, due to their connection to light availability (Fig. 4). Based on thousands of observations on hundreds of species (https://www.inaturalist.org/projects/mastif) we found that, for most species, seed production is concentrated in the upper crown that receives high sunlight. (We find exceptions to this, including in some of the genus Quercus and for predominantly tropical taxa that bear fruits directly on the stem.) Consistent with the fact that well-lighted trees produce more seeds than shaded individuals (43, 52, 62, 70), and light addition can increase branch reproductive effort (69), we further observed that edge trees concentrate reproductive effort on the side of the crown that is exposed to sunlight, and reproductive effort often decreases from the exposed outer crown to the shaded interior. This strong light dependence is consistent with the observation that many reproductive structures are photosynthetic (e.g., cones of the genus Pinus). The crown-surface area (CSA) model was evaluated as an alternative allometric model to Eq. 1 (SI Appendix, section S2).

Crown changes with diameter were evaluated from images obtained in open environments, supplemented with model analysis for effects of shading (SI Appendix, section S2). Tree height $H$ and base of the live crown $H_b$ provide estimates of the reproductive height $H_r = H - H_b$. A crown shape parameter describes how crown surface area $C_{\text{SA}}$ accumulates with height from the top of the tree (Fig. 4), a relationship that can change as crowns flatten with tree height. Because crown shape was not fully visible for crowded trees, we evaluated how declining reproductive height $H_r$ with shading from neighbors affected $C_{\text{SA}}$ in SI Appendix, section S2.

For completeness we implemented an alternative model, termed the ITD (91). This model uses size and crowding information from inventory data to define an individual’s crown status relative to neighbors. A principle difference from the CSA model is an assumption in ITD that crown shape is fixed; i.e., there is no flattening of the crown with tree size. We used the algorithm and parameter values from ref. 91 to evaluate tree height, crown shape, crown radius, and exposed crown depth. As with our model, $C_{\text{SA}}$ was obtained by integrating arc-wise and vertically overexposed crown depth (SI Appendix, section S2).

Fecundity Analysis. The analysis allows for increases in fecundity as trees mature and then increase in diameter, followed by change in the diameter–fecundity relationship as trees continue to grow. We estimated the number
of seeds produced by a tree f, which is proportional to reproductive mass through a constant (g per seed) $M_f = m \times f$ and, thus, affects only the intercept coefficient in Eq. 1. We use the MASTIF model for inference, a dynamic biophysical model for year-to-year and tree-to-tree seed production (44). This hierarchical Bayesian (state-space) model allows for conditional independence in crop-crowd and seed-trap data through latent states. A tree i is mature in year $t$, $z_{i,t}$, with probability $p_{i,t}$. Given that a tree is mature, $z_{i,t} = 1$, it produces seed with conditional fecundity $\psi_{i,t}$, which depends on tree size, shading, and climate. A mature tree remains mature $\{\psi_{i,t}(z_{i,t-1} = 1) = 1 \}$ and $\nu_{i,t}(z_{i,t-1} = 0) = 0$, although a mature tree may still have failed crops in some years. Expected fecundity for a tree year is $E(\psi_{i,t}) = p_{i,t} \psi_{i,t}$. Random effects on individual and year allow for wide variation between trees and over time that is typical of seed production (44). The posterior covariance between trees and years can take any form, avoiding assumptions of standard time-series models, important due to the quasi-periodic variation in time and varying levels of synchronicity between individual trees (44, 52).

To allow for size-based changes in the relationship with diameter, including potential declines, we use the model having expectation

$$E(\log f) = \log a + \log \psi$$

$$= \log a + \log b + \beta D + \beta D^2 + \ldots$$

where $\ldots$ refers to other variables that control the year-to-year variation in climate and the tree-to-tree differences related to competition from neighbors, local habitat, and species differences (SI Appendix, section 53 and Table 54). Note that this model includes a $D^2$ term that allows for changes related to size. Taken together, this form does not enforce the dynamic constraint on change in Eq. 2, instead allowing for exponentially increasing fecundity ($\beta_2 > 0$) as well as eventual declines at large size ($\beta_2 < 0$). Fecundity change with tree size does not necessarily follow a simple quadratic pattern on the log scale (Fig. 1) as specified conditionally in Eq. 3. Predictive distributions from this hierarchical state-space model can vary in shape due to the distribution of coefficients across predictors that include shading from neighbors and climate variables. Like most high-dimensional models, the posterior surface can be complex (118). In the MASTIF model (44) there will be combinations of maturation statuses, fecundity values, and dispersal distances that explain seed-trap dispersion better than other combinations. As the Markov chain Monte Carlo (MCMC) chains mix over different combinations of maturation status and conditional fecundity for each tree, the species-level $\beta$ parameters can find multiple modes. This behavior in the MCMC is, in fact, desirable, ensuring that the algorithm is indeed visiting the range of combinations of maturation/fecundity that each have differing capacity to explain the data. Predictive distributions from the fitted model display the range of behaviors that are consistent with the fitted models.
118. J. S. Clark, A. E. Gelfand, Hierarchical Modelling for the Environmental Sciences: Statistical Methods and Applications (Oxford University Press on Demand, 2006).